

Detection of the critically endangered catfish *Trichogenes claviger* (Siluriformes: Trichomycteridae) using environmental DNA



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Trichogenes corresponds to a basal taxon in family phylogeny, with three species in Atlantic Forest domain. One is *Trichogenes claviger*, occurring in forested streams in the headwaters of the Itapemirim basin, Espírito Santo. Until now, it was known to occur only at two localities in the Picada Comprida stream, a tributary of the rio Caxixe. The present study used environmental DNA (eDNA metabarcoding) for detection of this threatened with extinction species. Individuals were collected using two methods: dip nets for targeted capture and gillnets for passive sampling and generated barcode references for the species, which enabled the catfish detection through eDNA sampling at three of the ten investigated sites. In addition to the target species of this study, it was possible to detect other fish from the community associated with each sample site. A comparison of the methods for estimating species composition per sampled stream demonstrated that the eDNA metabarcoding approach identified their presence with more sensitivity than traditional collection methods, with fishing gears. The results also provide a better understanding of the spatial distribution dynamics of *T. claviger* and reinforce its conservation status. Thus, eDNA metabarcoding was demonstrated to be a powerful method for monitoring threatened and rare species.

Keywords: Atlantic Forest, Conservation, Espírito Santo, Streams, Threatened species.

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Trichogenes corresponde a um táxon basal na filogenia da família, com três espécies de distribuição endêmica aos riachos da Mata Atlântica. Uma delas, *Trichogenes claviger*, ocorre em riachos florestados de primeira e segunda ordem, nas cabeceiras do rio Itapemirim, ao sul do Espírito Santo. Até o momento, a espécie é conhecida de apenas duas localidades na sub-bacia do córrego Picada Comprida, um riacho formador do rio Caxixe. O presente estudo usou o DNA ambiental (eDNA metabarcoding) para detecção desta espécie de bagrinho ameaçada de extinção. Foram coletados indivíduos empregando técnicas tradicionais de amostragem e a partir destas amostras foram geradas referências barcode para a espécie, o que permitiu sua detecção através da amostragem por eDNA em três das dez localidades investigadas. Adicionalmente à espécie alvo deste estudo, foi também possível detectar outras espécies de peixes nas amostras de água. Uma comparação dos métodos de estimativa da composição das espécies por riacho amostrado indicou que a metodologia de eDNA metabarcoding é capaz de identificar com maior precisão a diversidade de espécies de peixes em um dado riacho em comparação com a metodologia tradicional com redes de pesca. Os resultados alcançados trazem um melhor entendimento da dinâmica de distribuição espacial de *Trichogenes claviger* podendo confirmar seu status de conservação. O poder de detecção do método de eDNA metabarcoding se revelou uma ferramenta para monitoramento de espécies raras ou ameaçadas.

Palavras chave: Conservação, Espécie ameaçada, Espírito Santo, Floresta Atlântica, Riachos.

INTRODUCTION

Trichomycteridae is one of the most diverse catfish families in South America, with more than 360 valid species (Ferraris, 2007; Rizzato, Bichuette, 2014; DoNascimento, 2015; Fricke *et al.*, 2024). Within this family lies the subfamily Trichogeninae, comprising a single genus, *Trichogenes*, whose species attracted great interest from the scientific community. Recognized as a relictual group within the family (*e.g.*, Grandcolas, Treweek, 2016), the sister-group position of the taxon (*sensu* Krell, Cranston, 2004) aggregates plesiomorphic features that are critical for polarizing characters in phylogenetic studies (Wiley, Lieberman, 2011; following the principles of Hennig, 1966).

The *Trichogenes* genus comprises three valid species endemic to the Atlantic Forest biome, being all the known species restricted to southwest Brazil (Fricke *et al.*, 2024). *Trichogenes longipinnis* Britski & Ortega, 1983, which is found in the headwaters of the rio Parati-Mirim and coastal rivers of southern Rio de Janeiro and northern São Paulo; *T. beagle* de Pinna, Reis & Britski, 2020, from the upper rio Doce in Minas Gerais and *T. claviger* de Pinna, Helmer, Britski & Nunes, 2010 from the rio Itapemirim headwaters in southern Espírito Santo.

The distribution of *Trichogenes claviger* is restricted to a single micro-basin. The species is known only from two proximate sites along the Picada Comprida stream, a headwater tributary of the Caxixe River, located in the municipality of Castelo, within the state

of Espírito Santo in southeastern Brazil. These small order rivers correspond to a well-preserved headwater environment. The fishes are small sized (largest known specimen reaching only 5.1 cm), being active swimmers in mid-water during the day and most of the time the only species of fish present in that stretch of the stream (Sarmento-Soares *et al.*, 2018). However, after ten years since its description, *T. claviger* remains poorly known geographically. Targeted field expeditions conducted between 2018 and 2023 in headwater streams of the Itapemirim and Jucu basins, including areas reported in Sarmento-Soares *et al.* (2018), failed to record new populations prior to the present study (field expeditions led by LMSS, RFMP, unpublished data).

Indeed, the restricted geographical area of occupation leads the species to be categorized as Critically Endangered (CR) in red lists (MMA, 2014; IUCN, 2022). *Trichogenes claviger* was ranked as a target species in the Brazilian National Plan for the Conservation of Endangered Atlantic Forest Fish and Aeglid crustaceans (PAN Peixes e Eglas da Mata Atlântica) (MMA/ICMBio, 2019), ended in December 2024.

These critically endangered populations face a high risk of extinction in the wild. Detecting their presence and understanding their distribution is particularly challenging with traditional, invasive methods such as gillnets and dip nets (Bonfil *et al.*, 2021), as these techniques can have a direct impact on the already vulnerable sampled community. Environmental DNA (eDNA) emerged as an important tool for assessing biodiversity and detecting rare and endangered species in their natural environment (Goldberg *et al.*, 2015; Thomsen, Willerslev, 2015; Cristescu, Hebert, 2018). The eDNA metabarcoding technique, originally developed in the late 1980s for the characterization of microbial communities from DNA recovered from soil samples, evolved into a powerful tool that enables detection of many organisms with great sensitivity. As this non-invasive approach allows the detection of multiple species through the DNA released into the environment, it is a particularly effective technique for finding traces of rare organisms (Thomsen, Willerslev, 2015; Boivin-Delisle *et al.*, 2020; De la Ossa-Guerra *et al.*, 2025). The data generated by this method are increasingly being used to direct and prioritize conservation efforts, from monitoring protected species (Biggs *et al.*, 2015) to informing environmental management policies (Lodge *et al.*, 2012; Goldberg *et al.*, 2016).

For eDNA metabarcoding to be able to detect a given species it is necessary that a reference sequence of a marker region is available for it, which is yet rarely the case for rich communities, especially at the Neotropics (Carvalho, Leal, 2023). Up to date, for the *Trichogenes* species, only *T. longipinnis* has partial reference sequences available for the mitochondrial regions D-loop, COI and Cytochrome B, more often used for barcoding applications, and has no references for the 12S rRNA region, targeted by the mini barcode marker *MiFish* (Takahashi *et al.*, 2023). Since this marker is the most widely used for fish eDNA metabarcoding, none of the *Trichogenes* species can be detected on conventional eDNA studies, they are yet invisible through the eDNA lenses.

A targeted multi-species approach was adopted for analyzing environmental DNA samples by eDNA metabarcoding using a generalist molecular marker for fish. Toward this end we validate the use of eDNA metabarcoding in the detection of *T. claviger* in streams close to points where the species had been previously reported, expanding records of the species' geographic distribution.

In this study we have generated new reference sequences from three *T. claviger* individuals for the 12S rRNA marker region, enabling its detection with eDNA metabarcoding. We also evaluate the eDNA metabarcoding and visual survey

methodologies complementarity, and applied both techniques to expand the known area of occurrence for the species, as a model for endangered species detection in remote areas.

MATERIAL AND METHODS

Study area. The study area is located between the municipalities of Castelo and Vargem Alta in mountainous areas of southern Espírito Santo. With approximately 85 km² of surface area and 45 km of perimeter, the area investigated included the headwaters of two river basins, the Itapemirim (formed by the Caxixe stream; Fig. 1A) and Jucu (formed by the rio Fruteiras; Fig. 1B). The elevation within the study area varies between 1,150 and 1,300 m. Two sampling expeditions were carried out: the first in September 2021 focused on specimen collection, and the second in March 2023 aimed at species detection using environmental DNA (eDNA) alongside traditional methods (namely, dip nets for targeted capture and gillnets for passive sampling). In the September expedition, for the investigation of the species *T. claviger*, it was necessary to collect some specimens to create a PCR and sequencing protocol using tissue fragments from these animals, so that, in the eDNA analyses, it would be possible to compare them with sequences available in the databases for species confirmation.

Ten georeferenced sites were sampled in the mountainous headwaters of the Itapemirim River and Jucu River basins, in streams located between Forno Grande State Park and the Private Natural Heritage Reserve (RPPN) Águia Branca (Figs. 1A, B). In the Itapemirim River basin, the sampled locations include three localities at ribeirão Braço Sul sub-basin: Point P2 at the Picada Comprida stream; Point P3 on a right-bank tributary of Picada Comprida stream; Points P1 and P8 on a right-bank tributary of ribeirão Braço Sul. In the ribeirão Caetés sub-basin: Points P4 and P5 on right-bank tributaries of ribeirão Caetés. In the córrego do Ouro sub-basin: Points P6 and P7 on left-bank tributaries of córrego do Ouro. In the rio Fruteiras sub-basin: Point P10 on a left-bank tributary of rio Fruteiras. Additionally, in the Jucu River basin, a single location was sampled: Point P9 on a right-bank tributary of the rio Jucu Braço Norte (Fig. 1).

Specimens collection for reference sequence generation. For field sampling, sieves and seines were used. At each point, a combination of fishing resources was used to ensure sampling without disturbing the local environment. The locality was sampled, whenever possible, by going through a stretch of approximately 50 m upstream. The specimens were euthanized with menthol, fixed in eugenol and transported to the laboratory, and cataloged. Some specimens collected were photographed in a field aquarium. Institutional abbreviations follow Sabaj (2025), with the Instituto Nacional da Mata Atlântica [formerly Museu de Biologia Professor Mello Leitão], Santa Teresa (MBML) as the institution of deposit of fishes sampled. Additionally, the genetic samples were processed at the Laboratório de Genética da Conservação (LGC) at Pontifícia Universidade Católica of Minas Gerais.

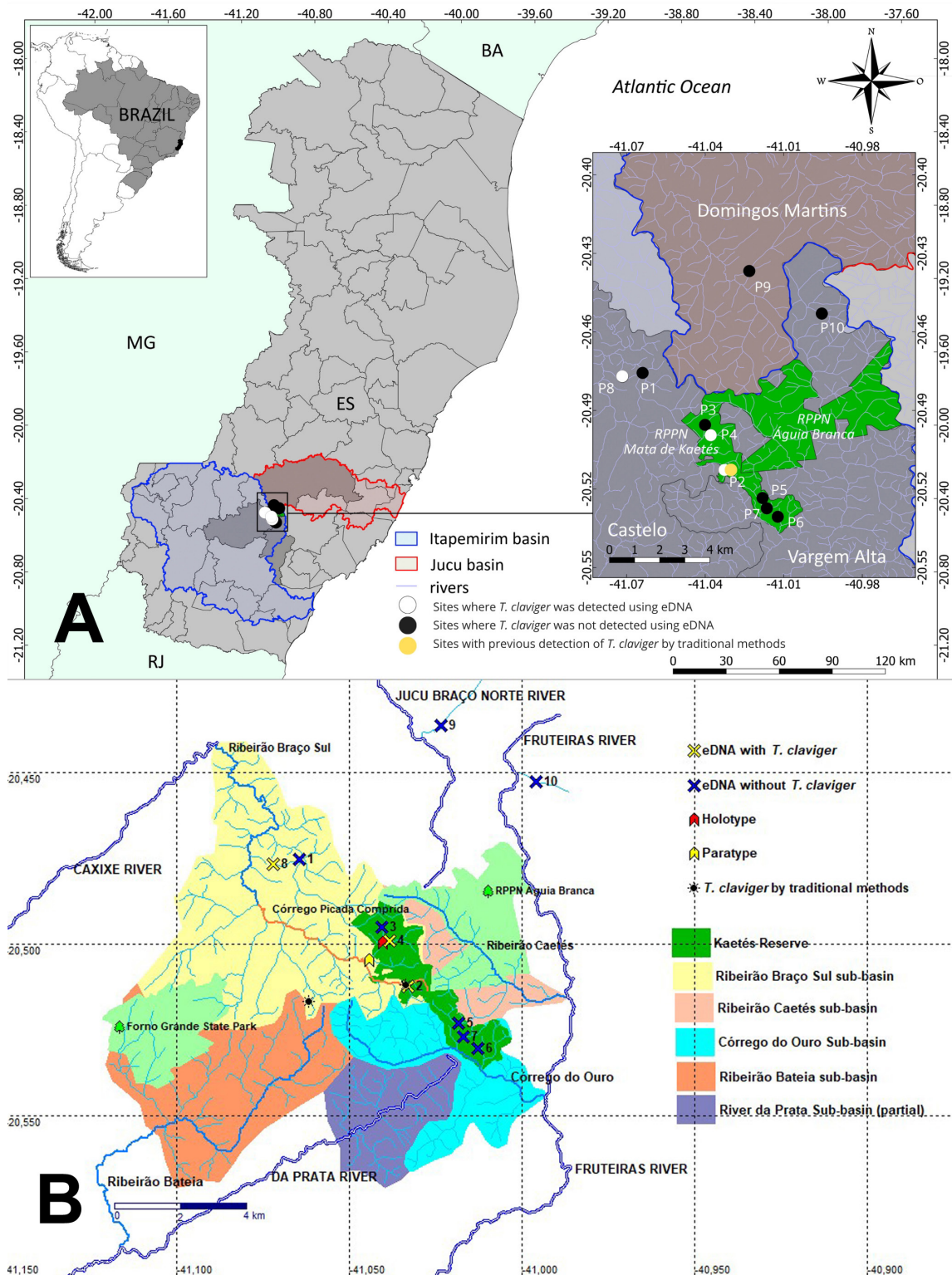


FIGURE 1 | A. Rio Itapemirim (blue contour) and rio Jucu (red contour) in southern Espírito Santo. The black circles indicate eDNA sampling points; white circles indicate detection of *Trichogenes claviger*. Areas highlighted in green, RPPN Mata de Kaetés and RPPN Águia Branca. **B.** Detail of detection samples on headwaters of Ribeirão braço Sul, the only river with records of *T. claviger* (in yellow contour). Detected presence (yellow X) in collecting localities 2, 4 (type locality) and 8. Absence (blue X) tentative in collecting localities 1, 3, 5, 6, 7, 9 and 10. Green contour corresponds to protected areas- Parque Estadual do Forno Grande, Reserva Particular do Patrimônio Natural, RPPN Águia Branca (in light green) and Reserva Particular do Patrimônio Natural, RPPN Mata de Kaetés (in dark green).

Generation of *Trichogenes claviger* 12S rRNA reference sequences. Five specimens of *T. claviger* were collected from the Braço Sul stream (sampling point P8) and RPPN Mata de Kaetés (P2), tributary of the rio Caxixe, Itapemirim basin (Fig. 2), for genetic reference construction. Tissue samples were obtained and deposited in the fish tissue collection of the Conservation Genetics Laboratory of PUC Minas. Genomic DNA was obtained by salt extraction adapted from Aljanabi, Martinez (1997). The 12S rRNA region was amplified using the primers MiFish-U-F 5'-GTCGGTAAACTCGTGCCAGC-3' (Miya *et al.*, 2015) and the Teleo REV ACTTCCGGTACACTTACCATG (Valentini *et al.*, 2016).



FIGURE 2 | Sampling of *Trichogenes claviger* during the eDNA fields. **A.** Live specimens in a field aquarium (Photo by J. P. Silva-Novelli). **B–C.** Collection in the RPPN Mata de Kaetés. **D.** Collection in Ribeirão Braço Sul.

The PCR program consisted of an initial cycle of denaturation at 97°C, followed by 35 cycles of denaturation at 95°C, girdling of the primers at 57°C and extension at 72°C. The success of the amplification was evaluated by electrophoresis on a 1% agarose gel. Samples that presented a sharp band with a size (pb) corresponding to the expected were selected for sequencing by the Sanger method for both fragment orientations forward (FWD) and reverse (REV). The quality of the sequences obtained from the FWD and REV amplicons was evaluated, and the sequences were subsequently merged to obtain the consensus sequence corresponding to the complete amplicon. The target 12S rRNA fragment spanned 650 bp and included the regions of three 12S rRNA markers: (1) NeoFish (~190 bp, Milan *et al.*, 2020); (2) MiFish (~170 bp, Miya *et al.*, 2015); and (3) Teleo (~70 bp, Valentini *et al.*, 2016), except for the MiFish FWD and Teleo REV annealing regions. These mini-barcodes are currently in use for most eDNA metabarcoding studies targeting ichthyofauna, as they can amplify a hypervariable region of the 12S rRNA gene, which contains information to effectively identify the tested fish to taxonomic family, genus, and species (Miya *et al.*, 2015). The newly generated 12S rRNA reference sequences for *T. claviger* are available in the NCBI nucleotide database under the accession numbers PX556613 to PX556616 (Tab. S1).

Water sampling and visual survey, and fish collections inventory. A total of 10 streams were sampled following the protocol established by Mendes *et al.* (2024a). At each site water was collected using a 1L sterile disposable bag, and filtered through 3 Sterivex (Millipore) cartridges, with 180 ml of water per cartridge with the aid of disposable syringes. A total of three cartridges (replicates) were used at each site. Two filtration controls with sterile mineral water were performed alongside field sampling to monitor potential contamination during the filtration process. All eDNA samples were collected before any other procedure, with the sites intact. Immediately after filtration, the filter cartridges were emptied and filled with 2 ml of ATL buffer (Qiagen) for sample preservation.

After collecting the water samples, a visual survey was performed. Each site was observed for 30 min, paying special attention to preferred niches: puddle water with slow currents, as well as sandy and leaf litter substrates. This method was chosen because *T. claviger* inhabits shallow, clear-water headwater streams with sandy substrates and accumulated plant material, where visual detection is feasible and traditional seining is often impractical due to irregular topography and limited space. Visual observation has been effectively used in previous ecological studies of the species (Sarmiento-Soares *et al.*, 2018). The list of species previously detected for the Itapemirim basin was obtained from Sarmiento-Soares *et al.* (2014), further referred to as fish collections, and used for comparison with eDNA metabarcoding and visual survey.

eDNA extraction, amplification and sequencing. The eDNA was extracted from the filters using a protocol developed by the Conservation Genetics Laboratory (LGC) (Mendes *et al.*, 2024b) with adjustments on the commercial DNeasy PowerWater Kit (Qiagen) such as initial step of DNA digestion with proteinase K and heating of elution buffer to 55°C. Purified extracts were checked with DNA concentration (ng/μL) and absorbance ratios (A260/280) estimated via a Nanodrop 2000 (Thermo, Tab. S2). The 12S rRNA MiFish marker was amplified with the

primers MiFish-U-F (5'-GTCGGTAAAACCTCGTGCCAGC-3') and MiFish-U-R (5'-CATAGTGGGGTATCTAATCCCAGTTTG-3') (Miya *et al.*, 2020), on a 25 µl final volume, using AmpliTaq Gold, with 2 µl DNA. The thermal cycles for this step were as follows: initial denaturation at 95°C for 2', followed by 35 cycles of denaturation at 95°C for 1', annealing at 60°C for 30', and elongation at 72°C for 1', followed by final elongation at 72°C for 7'. PCR negative controls were also included. PCR products were purified using Agencourt Ampure beads (Beckman Coulter) according to the manufacturer's instructions and used as templates for the second PCR, the library construction. This second PCR was conducted with Illumina adapters and indexes (Nextera Index kit®) following manufacturer protocol instructions. The library was quantified by real-time PCR using a KAPA Biosystems Quantification Kit, and sequenced on an iSeq100 (Illumina) using an iSeq v. 2 300 cycle (2x150 bp) kit. For each sampling site, all PCR replicates for the three sampling replicates were pooled together into a single library per site. The average raw output per sample was 64,502 read pairs. After bioinformatic processing, an average of 37,051 sequences per sample remained and were used for downstream analyses.

Bioinformatic and data analyses. Bioinformatics analyses were performed using a custom pipeline written in R v. 4.3.3 (R Development Core Team, 2024) based on Hilário *et al.* (2023). Primer and adapter removal was performed using the Cutadapt package (Martin, 2011). Low-quality reads filtering, reads merging into ASVs (Amplicon Sequence Variants) and chimeric sequences removal were performed using the DADA2 package (Callahan *et al.*, 2016). ASVs were grouped into OTUs using SWARM v. 2 (Mahé *et al.*, 2015) to evaluate the correlation between unique ASVs, OTUs and taxa on each sample.

All ASVs found in negative controls were completely removed from the other samples. All ASVs were taxonomically assigned using the BLASTr R package (Hilário *et al.*, 2024), by performing a parallelized BLASTn (Camacho *et al.*, 2009) against the NCBI core nucleotide database (Sayers *et al.*, 2022, *core_nt*, updated June/2025) combined with a custom database composed of 12S rRNA sequences for some species of the Brazilian fish fauna (Hilário *et al.*, 2023), along with the newly added reference sequences for *T. claviger* generated in this study. BLASTn searches were performed with identity and coverage cutoff values of 80%. All BLASTn matches with *identity* above 95% were evaluated for taxonomic assignment, revised by ichthyologists considering the local fauna, identity and coverage, and assigned to species, genus, or family taxa. All taxa with identity below 95% were discarded. Only taxa assigned to Actinopteri were kept on downstream analyses. Data analysis was performed using *dplyr* (Wickham *et al.*, 2023) and images were generated using *ggplot2* (Wickham, 2016). The Quantum-GIS v. 3.24.0 software was used to draw the map of geographic distribution.

Conservation action methods. To calculate AOO (Area of Occupation) and EOO (Extent of Occurrence), the Geospatial Conservation Assessment Tool (GeoCat, 2024), was employed regarding data on the occurrence of *Trichogenes claviger*. The EOO corresponds to the area contained in the smallest convex polygon including all the points of occurrence known for the species, while the AOO corresponds to the sum of the effective area of occurrence of the taxon within its EOO.

RESULTS

From the total 645,028 raw sequences obtained for all samples, 33.91% (218,761 sequences) were kept in bioinformatic processment, comprising 151 unique ASVs assigned to Actinopterii. Reads counts for each sample along the quality control and pipeline steps can be found on Tab. S3. All raw reads files and the samples table are available at <https://zenodo.org/> (DOI: 10.5281/zenodo.15230137). From these sequences, 8.76% (19,182 sequences) were assigned to *T. claviger*. This species was detected on the sampling sites P2, P4 and P8 using eDNA metabarcoding. The species was consistently detected across all three filtration replicates at each of these sites, supporting the reliability of the eDNA signal. It was the only species detected at these sites, except for the sampling site P8, where an ASV assigned to Pimelodidae was also detected. Concerning the visual survey, *T. claviger* was not detected at its original type locality, the sampling site P2, but it was spotted at the sites P4 and P8 (Fig. 3), showing the complementarity of these techniques.

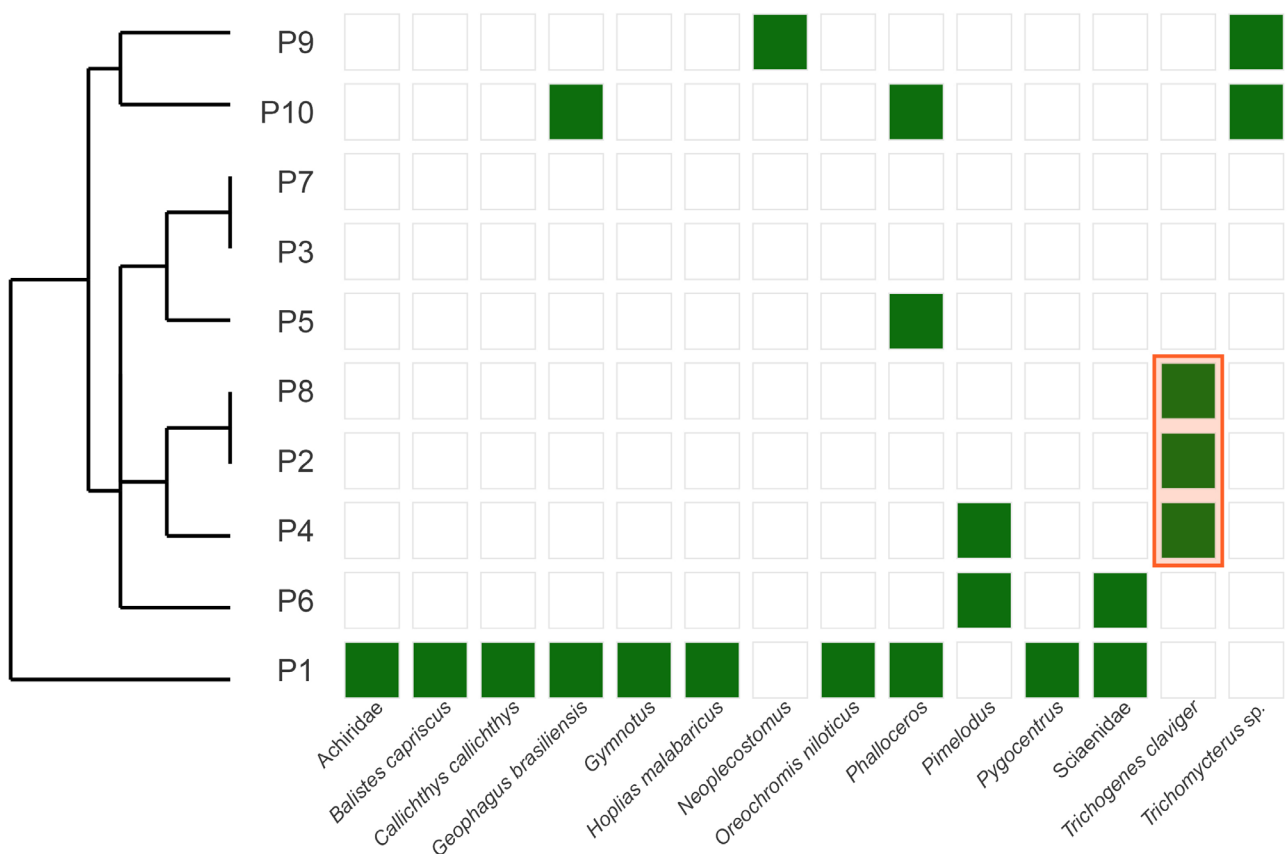


FIGURE 3 | Hierarchical clustering dendrogram of eDNA sampling sites based on fish species presence/absence data. The dendrogram was generated using Euclidean distance and Ward’s linkage method. *Trichogenes claviger* was detected only at sample sites P2, P4 and P8. Site P1 was the sample site with the highest number of species.

In addition to the species *T. claviger*, 14 other fish species were detected in water samples when considering all sampling sites (Tab. 1). All species detected via eDNA metabarcoding were represented by reference sequences in our curated 12S rRNA database, which included entries from the NCBI core nucleotide collection supplemented with a custom database for Brazilian freshwater fish (Hilário *et al.*, 2023) and the newly generated references for *T. claviger*. This ensured reliable taxonomic assignment for all reported detections. The species detected in this study belong to six orders, 10 families and 14 genera. The order Siluriformes was the most representative, comprising five

TABLE 1 | List of species detected by each sampling method: Fish collections, Visual Survey and eDNA metabarcoding. *Refer to fish species used as food source; # Refer to marine fish species.

ORDER/Family/Species	Sampling method		
	Fish collections	eDNA	Visual Survey
CHARACIFORMES			
Acestrorhamphidae			
<i>Astyanax microschemos</i> Bertaco & Lucena, 2006	X		
<i>Deuterodon giton</i> (Eigenmann, 1908)	X		
<i>Deuterodon janeiroensis</i> (Eigenmann, 1908)	X		
<i>Deuterodon parahybae</i> Eigenmann, 1908	X		
Crenuchidae			
<i>Characidium vidali</i> Travassos, 1967	X		
Erythrinidae			
<i>Hoplias malabaricus</i> (Bloch, 1794)*		X	
Serrasalminidae			
<i>Pygocentrus</i> sp.*		X	
CYPRINODONTIFORMES			
Poeciliidae			
<i>Phalloceros harpagos</i> Lucinda, 2008	X	X	
<i>Poecilia reticulata</i> Peters, 1859	X		
GYMNOTIFORMES			
Gymnotidae			
<i>Gymnotus carapo</i> Linnaeus, 1758	X	X	
PERCIFORMES			
Balistidae			
<i>Balistes capriscus</i> Gmelin, 1789**		X	
CICHLIFORMES			
Cichlidae			
<i>Geophagus brasiliensis</i> (Quoy & Gaimard, 1824)	X	X	
<i>Oreochromis niloticus</i> (Linnaeus, 1758)*		X	
Sciaenidae			
<i>Plagioscion</i> sp.**		X	
PLEURONECTIFORMES			
Achiridae			
<i>Hypoclinemus mentalis</i> (Günther, 1862)*#		X	
SILURIFORMES			
Callichthyidae			
<i>Callichthys callichthys</i> (Linnaeus, 1758)		X	
Loricariidae			
<i>Hypostomus affinis</i> (Steindachner, 1877)	X		
<i>Neoplecostomus microps</i> (Steindachner, 1877)	X	X	
Pimelodidae			
Pimelodidae sp.		X	
Heptapteridae			
<i>Rhamdia quelen</i> (Quoy & Gaimard, 1824)			X
Trichomycteridae			
<i>Trichogenes claviger</i> de Pinna, Helmer, Britski & Nunes, 2010	X	X	X
<i>Trichomycterus caudofasciatus</i> Alencar & Costa, 2004	X	X	

species, followed by two species of Perciformes, two of Cichliformes, two species of Characiformes and one species each of Cyprinodontiformes, Gymnotiformes and Pleuronectiformes. All species detected via eDNA metabarcoding were represented by reference sequences in our curated 12S rRNA database, which included entries from the NCBI core nucleotide collection supplemented with a custom database for Brazilian freshwater fish (Hilário *et al.*, 2023) and the newly generated references for *T. claviger*. This ensured reliable taxonomic assignment for all reported detections.

Sampling methods comparison. When we analyzed the three sampling methods, eDNA, visual survey and previous fish inventories even though they did not cover the same stretches, we observed that the estimates of species composition per sampled stream demonstrated that the fish collection method and eDNA complemented each other; considering that of the 12 species reported within the region on previous fish inventories, five were detected by eDNA metabarcoding (41.6%), and 2 fresh water species (40%) were detected only by eDNA (Tab. 1), namely Pimelodidae sp. and *Callichthys callichthys*. This demonstrates that eDNA metabarcoding should not be used as a replacement, and combining both strategies would maximize the comprehension of local ichthyofauna. Nonetheless, the fact that most previously known species were detected by eDNA metabarcoding with a small sampling effort highlights the power of this technique. The visual survey was only able to detect two species.

Updated *Trichogenes claviger* distribution. Our results expand the known distributional area of *T. claviger*. The species was previously known to occur only at its type locality at the Picada Comprida stream (Point P4; Fig. 1B), and one of its tributaries (non-type material collected with traditional methods marked as black dots in Fig. 1B).

Environmental DNA detected presence is assigned to Fig. 1B (yellow X) in the following three collecting localities P2, P4 and P8. Absence of detection of its presence signed in Fig. 1B (blue X). In this case, the tentative was not successful in collecting localities P1, P3, P5, P6, P7, P9 and P10 (see Fig. 1B). The results illustrated that *T. claviger* is still present in the region delimited for this study, and its detection in the samples expanded the distribution of the species to the ribeirão Braço Sul (Point 8). In calculating the Extension of Occurrence (EOO) that was previously shorter (IUCN, 2022) and now has 3,007 km² and a revised area of occupation (AOO) of 12 km² (Fig. 4), calculated by the GeoCAT system.

Conservation concerns. We identified an area of about 3 km² of EOO for the species. As *T. claviger* has an extremely restricted distribution, it still fits in the Critically Endangered (CR) category. The AOO was 12 km², which would classify the species as Endangered (EN). Despite the amplification of the AOO through eDNA detection, the species remains as Critically Endangered, deserving a very restrictive distributional area in mountainous slopes at southern Espírito Santo, Brazil. The few known records (< 5 locations) assert *T. claviger* as an endemic species of the Itapemirim basin. Although investigated, no records were achieved at the neighbor basin headwaters, at the rio Jucu. The results obtained corroborate the current threat status of the species (IUCN, 2022). Even with an increase in the locations of occurrence, the species continues to

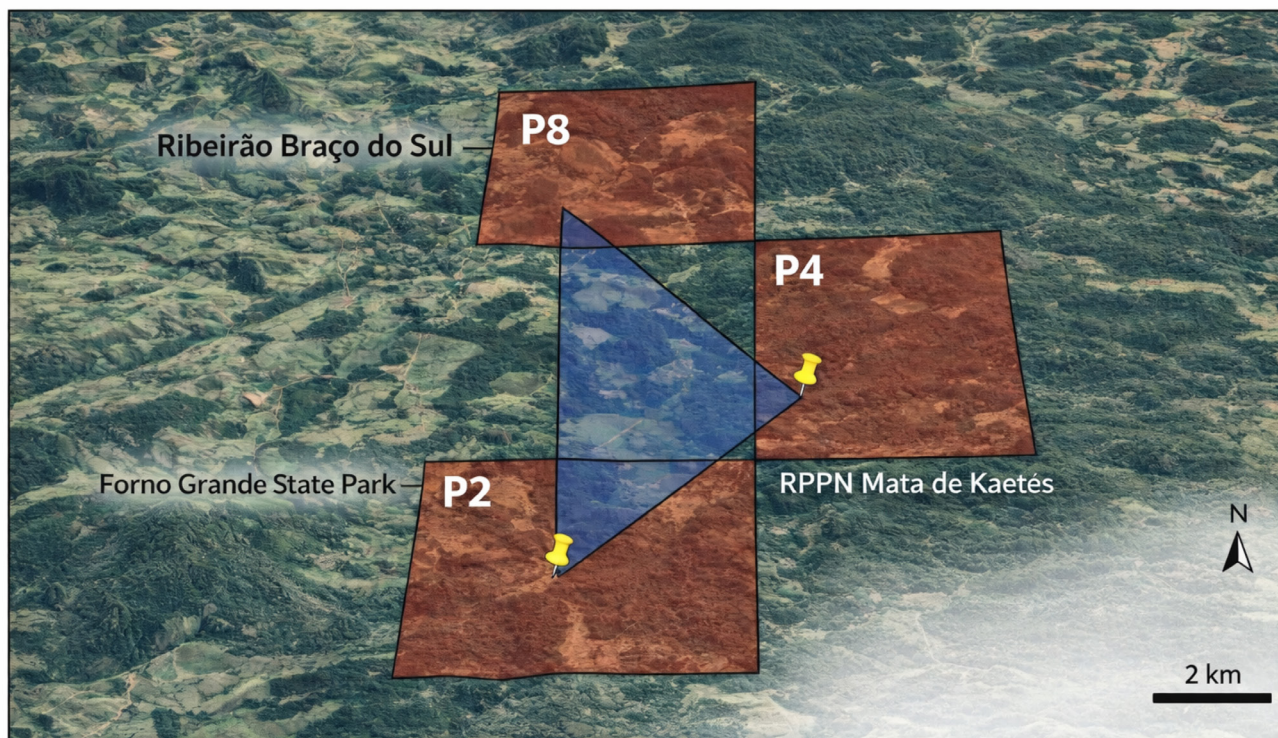


FIGURE 4 | Map of the result of the minimum convex polygon, with the GeoCAT, IUCN KEW assessment system, in order to categorize the risk of extinction of *Trichogenes claviger* AOO 12 km² and EOO 3,021 km². See text for further details. P2 at the Picada Comprida stream; P4 on right-bank tributaries of ribeirão Caetés; P8 on a right-bank tributary of ribeirão Braço Sul.

have environments restricted to the mountain foothills at southern Espírito Santo. The Point 8, located in ribeirão Braço Sul, in addition to being a short expansion of the distribution of the species, reveals that *Trichogenes claviger* is endemic to one single river sub-basin at upper rio Itapemirim.

DISCUSSION

Our target species, *Trichogenes claviger*, is known to inhabit small, high-altitude streams. These aquatic environments shelter small-sized fishes with limited dispersal potential, high endemism, and complex ecological interactions (Abilhoa *et al.*, 2011; Castro, Polaz, 2020). In low-order streams with clear water, shallow depth, and heterogeneous substrates (pebbles, submerged wood, sand, leaf litter), detecting local species using conventional methods is particularly challenging. How extensive is the species' distributional range? Can we predict it through alternative methods?

The successful detection of *T. claviger*, a rare and critically endangered species, using a multi-taxa eDNA metabarcoding approach underscores the sensitivity and applicability of this method for conservation monitoring. Environmental DNA has emerged as a non-destructive alternative to traditional, often destructive and labor-intensive freshwater monitoring methods (Cilleros *et al.*, 2019). Our methodological choices, including

filtering a substantial water volume (3×180 mL per site) and processing triplicate filtration replicates, were critical to increasing detection probability for low-abundance eDNA sources (Deiner *et al.*, 2015; Wood *et al.*, 2020). The consistent detection of *T. claviger* across all technical replicates at positive sites (P2, P4, P8) reinforces the robustness of the eDNA signal and aligns with recommendations for standardized protocols in rare species surveillance (Cooper *et al.*, 2022). This finding adds to a growing body of literature demonstrating the efficacy of eDNA metabarcoding for detecting elusive freshwater fishes, including other range-restricted and threatened species in diverse ecosystems (*e.g.*, Thomsen *et al.*, 2012b; Stoeckle *et al.*, 2015; Baker *et al.*, 2023).

The detection sensitivity of eDNA-based approaches has proven effective for monitoring endangered or elusive aquatic species, including freshwater fishes with restricted distributions (Thomsen *et al.*, 2012a; Miya *et al.*, 2015; Stoeckle *et al.*, 2015; Holmes *et al.*, 2024; Brys *et al.*, 2021; Baker *et al.*, 2023; Carvalho, Leal, 2023). Although eDNA detection probability is generally higher for species with larger populations, and can be influenced by body size and behavior (Thomsen *et al.*, 2012a; Pilliod *et al.*, 2013), our results demonstrate that eDNA metabarcoding is also a powerful tool for detecting rare, range-restricted species. This was corroborated by the congruence between eDNA signals and visual surveys where individuals were observed swimming. Moreover, eDNA succeeded in detecting the species at its type locality (site P4) where no individuals were sighted during our sampling, underscoring the method's sensitivity.

In the headwaters of the rio Itapemirim, eDNA facilitated the rapid detection of both endangered and allochthonous species, reducing sampling effort and minimizing impacts on these fragile forested streams (Deiner *et al.*, 2017; Schenekar *et al.*, 2020). The technique proved sensitive for tracking fish assemblages in these environments (Thomsen *et al.*, 2012b; Lacoursière-Roussel *et al.*, 2018), though detecting the target species in neighboring headwaters remains a task for future surveys.

The eDNA metabarcoding results revealed that *Trichogenes claviger* was frequently the only species detected at sites P2 and P4, and one of only two at P8. This low species richness aligns with the expected pattern for first- and second-order headwater streams in mountainous regions of the Atlantic Forest. Such environments are typically characterized by stringent physical conditions (*e.g.*, higher slope, less stable substrate, lower temperatures) and lower resource availability, resulting in naturally depauperate fish communities dominated by a few specialized species (Abilhoa *et al.*, 2011; Castro, Polaz, 2020). This phenomenon, described by the River Continuum Concept (Vannote *et al.*, 1980), is also observed in other coastal basins of the region, such as the upper Parati-Mirim River (habitat of *T. longipinnis*; Sazima, 2004) and the upper Doce River (where *T. beagle* occurs; Santos *et al.*, 2025), both of which harbor similarly low-diversity fish assemblages.

From an ecological perspective, the apparent dominance or exclusive occurrence of *T. claviger* in some reaches may reflect a combination of habitat specialization, such as the use of marginal microhabitats with leaf-litter accumulations, and reduced competition or predation pressure owing to the naturally species-poor community. However, this ecological simplicity also renders these populations extremely vulnerable. Habitat disturbances (*e.g.*, riparian deforestation, siltation, or introduction of non-native species) can have disproportionate impacts, because no more diverse fish assemblage exists to act as a functional buffer or a source for recolonization.

Our field observations further illustrate the specific habitat associations of *T. claviger*. The species was found in small riffles within forested areas above 1,000 m elevation in the upper ribeirão Braço Sul and rio Fruteiras headwaters (Itapemirim basin). At site P2 (Córrego Picada Comprida, RPPN Mata de Kaetés), individuals were observed swimming near the margin under the shade of large *Eucalyptus* trees, approximately 50 m from plantations (elevation 1,145 m). The fast-flowing, transparent water at sites P4 (type locality) and P8 (Ribeirão Braço Sul) allowed visual confirmation of actively swimming individuals, with site P8 appearing to function as a nursery, hosting dozens of juveniles and sub-adults. Notably, at all sites where *T. claviger* was detected, it was the only fish recorded in the immediate environment and was a common component of the assemblage. This pattern of isolated headwater populations mirrors that reported for congeners *T. longipinnis* (Sazima, 2004) and *T. beagle* (Santos *et al.*, 2025).

The rapid advancement of genomic technologies is promising. eDNA not only aided in detecting *T. claviger* but also provides a framework for continuous monitoring of a species previously known only from its type locality. The 12S rRNA marker reference developed here for *T. claviger* proved effective, highlighting the importance of generating reference sequences, especially for endangered species inhabiting sensitive environments, to unlock their detection via eDNA metabarcoding.

Some species detected by eDNA metabarcoding were unexpected in freshwater ecosystems, belonging instead to marine habitats. All of these marine taxa are commonly used as food sources and are commercially available (Fig. 3). Importantly, they were detected exclusively at site P1, the most anthropogenically disturbed site, located close to human settlements. This finding underscores the high sensitivity of the eDNA approach and its potential to trace human-mediated subsidies or contamination in aquatic systems.

In this context, our results reinforce the need for headwater-specific conservation strategies. Protecting these environments goes beyond preserving individually threatened species; it is about safeguarding unique ecological processes and simplified yet functionally specialized communities that are irreplaceable for the integrity of the river continuum. eDNA-based monitoring programs, as demonstrated here, are promising tools for tracking not only the presence of target species but also compositional changes in these sensitive communities over time. Our results may be used by Brazilian environmental agencies, in partnership with researchers and local communities, to prioritize and implement timely conservation actions that ensure the protection of this delicate environment and its endemic species.

Material examined. Brazil. *Trichogenes claviger*. Espírito Santo State, rio Itapemirim. MBML 3289, holotype, 50.4 mm SL, MBML 3290, 11 paratypes, 43.2–14.8 mm SL, tributary of córrego Picada Comprida. MBML 3987, 13, 36.7–23.7 mm SL, córrego Picada Comprida. MZUSP 124882, 33 (2 cleared and stained), 49.9–25.5 mm SL, tributary of córrego Picada Comprida. MBML 14089, 1, 31.4 mm SL, MBML 14091, 8, 48.8–30.2 mm SL, córrego Picada Comprida. MBML 14206, 10, 38.9–29.2 mm SL, MBML 14207, 3, 33.7–30.1 mm SL, tributary of rio Braço Sul.

Genetic samples (Tissue-TS). Brazil, Espírito Santo State, rio Itapemirim. LGC 8125 (TS); LGC 8126 (TS); LGC 8127 (TS); LGC 8128 (TS); LGC 8129 (TS).

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AUTHORS' CONTRIBUTION

Juliana Paulo da Silva-Novelli: Conceptualization, Data curation, Formal analysis, Funding acquisition, Investigation, Methodology, Project administration.

Heron Oliveira Hilário: Conceptualization, Data curation, Formal analysis, Investigation, Methodology, Resources, Software, Writing–original draft.

Daniel Cardoso Carvalho: Conceptualization, Data curation, Formal analysis, Funding acquisition, Investigation, Methodology, Resources, Software, Validation, Writing–original draft.

Ronaldo Fernando Martins-Pinheiro: Formal analysis, Investigation, Methodology, Software, Validation, Visualization, Writing–original draft, Writing–review and editing.

Luisa Maria Sarmiento-Soares: Conceptualization, Data curation, Formal analysis, Investigation, Methodology, Project administration, Supervision, Validation, Visualization, Writing–original draft, Writing–review and editing.

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ETHICAL STATEMENT

Sistema de Autorização e Informação sobre Biodiversidade (SISBIO) for the collection license (registration 79046–1/2).

DATA AVAILABILITY STATEMENT

The authors confirm that the data supporting the findings of this study are available within the article.

AI STATEMENT

The authors did not use any AI-assisted technologies in the creation of this manuscript or its figures.

COMPETING INTERESTS

The authors declare no competing interests.

SUPPLEMENTARY MATERIAL

Supplementary Material File

PEER REVIEW

Peer Review File

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